Using L-Arginine-Functionalized Gold Nanorods for Visible Detection of Mercury(II) Ions

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Abstract: A rapid and simple approach for visible determination of mercury ions (Hg^{2+}) in aqueous solutions was developed based on surface plasmon resonance phenomenon using L-arginine-functionalized gold nanorods (AuNRs). At pH greater than 9, the deprotonated amine group of L-arginine on the AuNRs bound with Hg²⁺ leading to the side-by-side assembly of AuNRs, which was verified by transmission electron microscopy images. Thus, when Hg²⁺ was present in the test solution, a blue shift of the typical longitudinal plasmon band of the AuNRs was observed in the ultra violet-visible-near infrared (UV-Vis-NIR) spectra, along with a change in the color of the solution, which occurred within 5 min. After carefully optimizing the potential factors affecting the performance, the L-arginine/AuNRs sensing system was found to be highly sensitive to Hg²⁺, with the limit of detection of 5 nM (S/N = 3); it is also very selective and free of interference from 10 other metal ions (Ba²⁺, Ca²⁺, Cd²⁺, Co²⁺, Cs⁺, Cu²⁺, K⁺, Li⁺, Ni²⁺, Pb²⁺). The result suggests that the L-arginine-functionalized AuNRs can potentially serve as a rapid, sensitive, and easy-to-use colorimetric biosensor useful for determining Hg²⁺ in food and environmental samples.

Keywords: colorimetric sensor, food safety, green synthesis, heavy metal, surface plasmon resonance, water quality

Practical Application: In this paper we present a rapid, sensitive, and easy-to-use method for detecting amount of mercury ions present in water. Using L-arginine-functionalized gold nanorods, which change in color when mercury is present in the sample, we are able to detect as low as 5 nM of mercury within 5 min. This method can be useful for routine measurement of mercury present in food and environmental samples

Introduction

Mercury is one of the most toxic elements that have adverse effects on humans and biological system (Tchounwou and others 2003; Selin 2009), and the water-soluble mercury(II) ion (Hg²⁺) is one of the most common forms of mercury pollution. Mercury tends to be bioaccumulated and biomagnified in aquatic food chains (Clarkson and others 2003; Leopold and others 2010). Consequently, consumption of fish and other seafood is one of the most important exposure routes for mercury (Al-Mughairi and others 2013; Greenfield and others 2013). Based on the Natl. Health and Nutrition Examination Survey, mercury levels were 4-fold higher in people who ate 3 or more servings of fish per month than in those who did not consume seafood (Schober and others 2003). Since fish contain higher levels of mercury that may harm an unborn baby or young child's developing nervous system, the U.S. Food and Drug Administration and Environmental Protection Agency have advised that pregnant women, nursing mothers, and young children completely avoid certain fish and other seafood (especially shark, swordfish, king mackerel, and tilefish) that contain significantly high levels of mercury (FDA 2014). Mercury exposure can also damage a variety of human organs, such as kidney and brain and compromise immune, nervous, and endocrine systems (Clarkson and others 2003; Nolan and Lippard 2008; de Souza and de Carvalho 2009; Mutter and others 2010; Mamdani and Vettese 2013).

The determination of mercury ions in contaminated samples is generally performed by instrumental methods, such as inductively coupled plasma mass spectrometry (Ugo and others 2001), high-performance liquid chromatography (Balarama Krishna and others 2007), atomic absorption spectrometry (Hsu and others 2011), and electrochemical methods (Zhu and others 2009). Although these methods offer excellent sensitivity, their major drawbacks are high cost and complex and time-consuming test protocols. Hence, the development of a simple, rapid, economical, and sensitive method for mercury detection is needed for detecting mercury ions in food and environmental samples (Male and others 2013; Mansour 2014; Yasri and others 2014).

Colorimetric sensors are convenient and attractive because they allow prompt monitoring without the need for read-out devices or trained personnel to perform the test (Nolan and Lippard 2008; Lim and others 2012). Much research has been done on designing optical sensing systems for detecting mercury ions based on conjugated polymers (Qu and others 2014), biomolecules, such as proteins (Guo and others 2011), DNA (Xue and others 2008; Wei and others 2014), oligonucleotide (Liu and others 2007), and quantum dots (Li and others 2011; Chen and others 2014), and inorganic materials (Tao and others 2014). Moreover, functionalized gold nanoparticles (AuNPs) have been widely used in biosensors due to the distinct surface plasmon resonance properties (Jans and Huo 2012; Wang and Gunasekaran 2012; Hutter and Maysinger 2013; Wang and others 2015), including for the detection of heavy metals, such as Hg²⁺, Pb²⁺, Cu²⁺ (Lin and others 2011; Chen and others 2014). Compared to AuNPs, which are typically spherical and exhibit 1 plasmon band at around 520 nm, anisotropic nanoparticles such as nanorods exhibit 2 plasmon bands, owing to the oscillation of electrons along transversal

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and longitudinal directions. The transversal peak of gold nanorods (AuNRs) remains at around 500 nm, however, their longitudinal plasmon band is located anywhere in the visible to near infrared range (600 to 900 nm), depending on their aspect ratio (Jana and others 2001). Interestingly, the longitudinal plasmon band of AuNRs has a higher extinction coefficient, rendering AuNRs more sensitive to changes in the dielectric properties of the surrounding environment, which could improve the performance of biosensors (Vigderman and others 2012; Placido and others 2013). Herein, we describe a green method using L-arginine, one of the proteinogenic amino acids, to modify AuNRs and use this greenfunctionalized AuNRs for visible detection of Hg^{2+} in water. The possible mechanism for this visible detection system and its sensitivity, selectivity, limit of detection (LOD) of Hg^{2+} are also discussed.

Materials and Methods

Materials

Hydrogen tetrachloroaurate (III) trihydrate (HAuCl₄·3H₂O), sodium borohydride (NaBH₄, \geq 98%), silver nitrate (AgNO₃, 99.85%) were from Acros Organics (Morris Plains, NJ, USA). Cetyltrimethyl ammonium bromide (CTAB) was from MP Biomedicals (Solon, OH, USA). L-ascorbic acid (AA, \geq 99%) and sodium hydroxide (NaOH) were from Fisher Scientific.

For heavy metals, Mercury(II) nitrate monohydrate $(Hg(NO_3)_2.H_2O, \geq 98\%)$, lead(II) chloride (PbCl₂, 99%), nickel(II) chloride hexahydrate, (NiCl₂·6H₂O, 97%), zinc chloride, (ZnCl₂, 97%), lithium chloride, (LiCl, \geq 99%) were from Acros Organics. Cupric chloride, dihydrate, (CuCl₂·2H₂O, 99%) was from Sigma-Aldrich (St. Louis, MO). Calcium chloride



Figure 1–Scheme for seed-mediated method for synthesis of gold nanorods (AuNRs). The seed is first prepared by mixing cetyltrimethyl ammonium bromide (CTAB) with HAuCl₄, followed by adding NaBH₄ and kept at 29 °C for 2 h. Then, the seed solution is added to the growth solution, which includes HAuCl₄, CTAB, HCl, and AgNO₃ and L-ascorbic acid, and kept at 29 °C overnight to complete the synthesis of AuNRs.



Figure 2–Effect of adding 1 mM Hg²⁺ to L-arginine-functionalized gold nanorods (AuNRs) solution: color (a) before (left) and after (right) Hg²⁺ addition; absorption spectra (b); and TEM micrographs before (c) and after (d) Hg²⁺ addition.

dihydrate (CaCl₂·2H₂O, \geq 99%), barium dichloride dihydrate (BaCl₂·2H₂O, 100% w/v), potassium chloride (KCl, \geq 99%), from Fisher Scientific (Rockford, IL). Cobalt chloride hexahydrate (CoCl₂·6H₂O, \geq 99%) and cadmium chloride (CdCl₂, \geq 99%) were purchased from MP Biomedicals. These chemicals were used as received without any purification. All aqueous solutions were prepared in deionized water of resistivity \geq 18.2 M Ω ·cm (Ultrapure water system, Millipore, (Billerica, MA)).

Equipment

Absorption spectra (400 to 1100 nm) were recorded using Lambda 25 UV-Vis-NIR spectrophotometer (PerkinElmer



Figure 3–(a) Color of L-arginine-functionalized gold nanorods solution with 1 mM Hg²⁺ at different pHs. The pH of solution is controlled by adding different amount of 0.1 M NaOH. (b) Chemical structure of L-arginine, and (c) structure of L-arginine at pH higher than 9 (pKa₂) but lower than 12.48 (pKa₃).

(Perkin Elmer, MA)). Morphology of AuNRs was determined using FEI (Hillsboro, OR) Tecnai T12 transmission electron microscope (TEM) operating at 120 kV. Typically, samples were prepared by dropping the solution on a carbon-coated 400 mesh copper grid and excess solution was removed by wicking with filter paper. The grid was allowed to dry at room temperature before imaging.

Synthesis and L-arginine functionalization of AuNRs

AuNRs were synthesized following a published seed-mediated protocol with slight modification (Orendorff and Murphy 2006). Figure 1 shows the scheme for synthesis of AuNRs. Briefly, a seed solution is first prepared by reducing 250 μ L 10 mM HAuCl₄ in 9.75 mL 0.1 M CTAB with 800 μ L ice-cold 0.01 M NaBH₄ under vigorous stirring for 1 min and placing in water bath at 29 °C for 2 h. A yellow to brownish-yellow color change of the solution signifies the formation of gold seeds.

The growth solution was prepared by adding 1 mL 10 mM HAuCl₄ into 40 mL of 0.1 M CTAB with gentle mixing, followed by 40 μ L of 0.1 M AgNO₃ and 800 μ L of 1 M HCl, and finally 160 μ L of 0.1 M L-AA. The color of the solution eventually changes from dark yellow to being colorless. For synthesizing AuNRs, 160 μ L of the seed solution was added into the growth solution and kept overnight in a water bath at 29 °C. The samples were characterized by UV-Vis-NIR spectroscopy and TEM.

For L-arginine functionalization, 1 mL of the AuNRs solution was gently mixed with 1 mL 0.5 mg/mL L-arginine. After 5 min, the pH of the mixed solution was adjusted by successive addition of 0.1 M HCl/NaOH to determine the optimal pH. The resulting solution was incubated for 5 min at room temperature and its absorption spectrum was recorded. The morphology of AuNRs was determined by TEM.

Determination of Hg²⁺

To determine the Hg^{2+} content in water, 1 mL of different known concentrations of Hg^{2+} were added into the Larginine/AuNRs solution and incubated at room temperature for 5 min. The observed color and absorption spectra of the solution were recorded and the morphology of AuNRs were determined by TEM. To evaluate the specificity of Hg^{2+} detection, 10 mM solution of different metal ions, such as Ba^{2+} , Ca^{2+} , Cd^{2+} , Co^{2+} ,



Figure 4–Mechanism of using L-arginine-functionalized AuNRs for visible detection of Hg^{2+} ions. First, negatively charged carboxylic group of L-arginine binds to positively charged surface of the AuNRs. Then, at pH > 9.0, in the presence of Hg^{2+} ions, the L-arginine/AuNRs assemble side-by-side, which leads to red to green color change.

Cs⁺, Cu²⁺, K⁺, Li⁺, Ni²⁺, Pb²⁺, and Zn²⁺ were added into the L-arginine/AuNRs solution in the presence of 1 mM of Hg^{2+} .

Results and Discussions

Experimental optimization and proposed mechanism

Sensing strategy. The color of the L-arginine/AuNRs solution (Figure 2a) changed from red, which indicates the presence of AuNRs in the solution, to green with the addition of Hg^{2+} (1 mM). When the same experiments were performed using AuNRs without L-arginine functionalization, the solution color remained red with or without the addition of 1 mM of Hg^{2+} . Hence, the L-arginine functionalization of AuNRs is essential to produce color change and visible detection of Hg^{2+} . This color change occurs and becomes visible to the bare eye within 5 min of Hg^{2+} addition.

Generally, color change in solutions containing AuNPs occurs due to the shift in the plasmon band. Typically AuNRs exhibit 2 plasmon bands (Figure 2b). The more prominent peak in the NIR region corresponds to the longitudinal plasmon band, while the smaller peak around 500 nm is associated with the transverse plasmon band (Fang and others 2012). After Hg^{2+} addition there is a blue shift, that is, shift toward shorter wavelengths, in both the plasmon bands; however, the shift in the longitudinal plasmon band is more substantial along with significant decrease in signal intensity compared to that of the transverse plasmon band (Figure 2b). This suggests that, as expected, the longitudinal plasmon band is more sensitive to changes in the dielectric properties of the surrounding environment (Wang and others 2010).

This blue shift with decrease in amplitude of longitudinal surface plasmon peak might result from the modest side-by-side assembly of AuNRs (Wang and others 2010). TEM images of the

L-arginine/AuNRs show that the AuNRs were well-dispersed within the solution before adding Hg^{2+} (Figure 2c); however, after Hg^{2+} addition, the AuNRs seemed to assemble along their longitudinal axis to some extent (Figure 2d). Side-by-side assembled structure is considered responsible for producing a blue shift of the longitudinal surface plasmon peak and concomitant color change (Xu and others 2011).

Effect of pH. Under the same experimental conditions, the system color change is dependent on the solution pH (Figure 3a). Over the tested pH range of 2.77 to 11.96, the color change occurs only when pH is higher than 9.0. This is due to the effect of pH on L-arginine. Figure 3b shows the structure of L-arginine, which has 3 pKa values: $pKa_1 = 2.02$ (α -carboxylic acid), $pKa_2 = 9.0$ (amine group), and $pKa_3 = 12.48$ (side chain group of amino acid) (Kumler and Eiler 1943). As can be seen from Figure 3c, when pH value was higher than 9.0 (pKa_2), the color started to change in the presence of Hg²⁺. At this pH, both the carboxylic acid and amine groups of L-arginine are deprotonated (Figure 3c), which might become a bridge between Hg²⁺ and AuNRs in the sensing system.

During the synthesis of AuNRs, CTAB, a well-known cationic surfactant, forms a positively charged CTAB bilayer on the AuNRs (Gole and others 2004). Therefore, the attractive electrostatic interactions between the negatively charged carboxylic acid of L-arginine and positively charged CTAB bilayer on AuNRs might primarily be responsible for forming stable L-Arginine/AuNRs, while pH value is greater than 2.02 (pKa₁) (Gole and others 2004). When the pH value increase to more than 9, the deprotonated amine group of L-arginine would bind to Hg²⁺, which is a possible reason for the AuNRs to assemble side-by-side. This suggested





AuNRs assembly mechanism is illustrated in Figure 4. We chose 9.5 as the optimal pH value for Hg^{2+} detection using our system.

Measuring Hg²⁺

Calibration data for the determination of Hg²⁺ were obtained under an optimized set of experimental conditions. As seen in Figure 5a, the blue shift of the longitudinal band increased with increasing Hg²⁺ concentration. However, there appears to be 2 distinct ranges of Hg²⁺ concentration (C, μ M) over which the peak shift increases fairly linearly: 1 to 20 μ M and 20 to 600 μ M. The corresponding linear regressions (each with $R^2 = 0.99$) are as follows:

1 to 20 μ M Hg²⁺: Blue Shift (nm) = 0.5307C + 10.464 20 to 600 μ M Hg²⁺: Blue Shift (nm) = 0.0416C + 22.223.

The LOD for Hg^{2+} , at a signal-to-noise ratio of 3, is calculated at 5 nM (MacNaught 2000). Figure 5b shows that in the presence of Hg^{2+} , the solution color changes from red to green, with the intensity of color change becoming deeper with increasing Hg^{2+} concentration.

Selectivity test

The selectivity of Hg^{2+} detection of our sensor in the presence of several other metal ions (Ba^{2+} , Ca^{2+} , Cd^{2+} , Co^{2+} , Cs^+ , Cu^{2+} , K^+ , Li^+ , Ni^{2+} , Pb^{2+} , and Zn^{2+}) at 10 times the concentration of Hg^{2+} was very good. As shown in Figure 6a, the red-to-green color change occurred only for Hg^{2+} . The ratio of absorbance values at 727 nm and 796 nm, which are wavelengths of the longitudinal bands before and after the blue shift after adding Hg^{2+} , was used to determine the selectivity of Hg^{2+} . Further, to reduce the influence of other factors, the 727 nm/796 nm absorbance ratio value of blank (that is, water) was subtracted from that of the metal solutions (Figure 6b). As can be seen from Figure 6, none of the metal ions tested interfered with the detection of Hg^{2+} even at 10 times the concentration. The small but measurable signals obtained with other metal ions (Figure 6b) are considered to be the result of changing dielectric properties of those metal ions at high concentrations, as they were tested at 10 times the concentration of Hg^{2+} . Further, the system color changed only in the presence of Hg^{2+} (Figure 6a). One possible explanation is that at pH > 9, the deprotonated amine group on L-arginine can bind only to Hg^{2+} (Palecek and others 1982; Delnomdedieu and others 1989; Safavi and Farjami 2011).

Conclusions

A simple, fast, and easy-to-use colorimetric assay for selective and sensitive determination of Hg^{2+} in water was developed based on surface plasmon resonance of L-arginine-functionalized AuNRs. By measuring the specific longitudinal plasmon band shift, which might relate to the side-by-side assembly of AuNRs, the L-agrinine/AuNRs system can be used to measure Hg^{2+} as low as 5 nM. The selectivity of Hg^{2+} detection is high as the system color change is largely unaffected by the presence of other metal ions (Ba²⁺, Ca²⁺, Cd²⁺, Co²⁺, Cs⁺, Cu²⁺, K⁺, Li⁺, Ni²⁺, Pb²⁺) even at 10 times the concentration of Hg^{2+} . This sensing method uses an environmentally friendly approach to functionalize AuNRs and provides a cost-effective way for detecting highly toxic Hg^{2+} in food and environmental samples.

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